# Particulate matter reduction in Eindhoven



R. VERVOORT ir. r.p.m.vervoort@tue.nl a



B. BLOCKEN prof. dr. ir. b.j.e.blocken@tue.nl <sup>a, b</sup>



T. VAN HOOFF dr. ir. t.a.j.v.hooff@tue.nl b, a



R. GIJSBERS dr. ir. r.gijsbers@enstechnology.nl <sup>c</sup>

High traffic intensities in cities are a cause of high particulate matter (PM) concentrations due to tailpipe emissions and tire and brake wear. The objective of this study, is to assess whether PM removal in parking garages by implementation of electrostatic precipitation, can reduce PM concentrations in the wider urban surrounding.

**Keywords:** Air quality; Air pollution; Fine dust; Particulate matter; Computational fluid dynamics (CFD); Electrostatic precipitation; Positive ionization; Urban physics.

# The effects of particulate matter exposure

The collection of all solid and liquid particles suspended in the atmosphere, known as particulate matter (PM), is currently one of the most dangerous forms of air pollution. According to the World Health Organization (WHO), daily and long-term exposure to PM is strongly related to human morbidity and mortality [1]. Health effects are closely related to the size of the inhaled particles [1]; large particles can be filtered by the nose and throat, however, particles smaller than 10  $\mu m$  (PM $_{10}$ ) can enter the bronchi and lungs. Particles smaller than 0.1  $\mu m$ 

<sup>&</sup>lt;sup>a</sup> Building Physics and Services, Department of the Built Environment, Eindhoven University of Technology, P.O. box 513, 5600 MB Eindhoven, The Netherlands

<sup>&</sup>lt;sup>b</sup> Building Physics Section, Department of Civil Engineering, KU Leuven, Kasteelpark Arenberg 40, bus 2447, 3001 Leuven, Belgium

<sup>&</sup>lt;sup>c</sup> ENS Technology, Grotestraat 41, 5431 DH Cuijk, The Netherlands.

(PM<sub>0.1</sub>) can directly be transported into the bloodstream [1]. Many studies have linked PM exposure to lung cancer, respiratory, cardiovascular and cardiopulmonary diseases (e.g. [2-3]) (an extensive overview of references is given in the paper by Blocken et al. [3]). Furthermore, links with Alzheimer, Parkinson, dementia, multiple sclerosis and stroke incidences are found (e.g. [2-4]). Groups with pre-existing lung or heart diseases, elderly people and children are especially vulnerable [2]. PM exposure affects lung development of children; i.e. lung functioning is impaired and lung growth rate is chronically reduced [5]. According to the Organization for Economic Cooperation and Development (OECD), air pollution will become the world's top environmental cause of premature mortality by 2050 when no measures are taken [6]. According to the European Environment Agency (EEA), the world-wide number of premature deaths due to PM exposure, as presented in Figure 1, will climb from approximately 1.5 million today, to 3.5 million in 2050 [7]. Given that good health and a long life are highly valued in society, analysis shows that economic costs of air pollution are significant [8].

# Air pollution in the city

The WHO provides limits for the annual mean PM concentrations. i.e. for  $PM_{10}$  the annual mean is set at  $20 \mu g/m^3$  and the 24 hour mean is set at  $50 \mu g/m^3$ 

[9]. For PM<sub>2.5</sub> these values are set at 10 and 25 μg/m<sup>3</sup> respectively [9]. Worldwide, more than 80% of the people that live in urban areas are exposed to air quality levels that exceed the WHO limits [10]. While in Asia all the global cities fail to meet the 10 μg/m<sup>3</sup> PM<sub>2.5</sub> limit, in Europe few are able to pass (e.g. Helsinki, Edinburgh and Stockholm) [10]. In a global context, ambient PM<sub>2.5</sub> concentrations in European cities seem relatively low. For example, when considering PM<sub>2.5</sub> concentrations, annual mean concentrations in Paris are exceeded by a factor 1.8 while in Delhi this is a factor 12.2 [10]. Although the WHO limits are defined strictly, it is important to keep in mind that epidemiological studies were unable to define a specific threshold for which PM concentrations have no effect on human health [2]. It is expected that there is a very wide range in susceptibility; some individuals might be at risk when exposed to very low concentrations [2].

High ambient concentrations of PM are found in regions with high traffic intensity, including urban areas, due to tailpipe emissions and tire and brake wear [11]. Other studies showed that PM is accumulated in parking garages [12–13]. In the surroundings of parking garages the ambient PM concentrations are highly dependent on parameters such as traffic intensity, meteorological conditions and urban geometry.

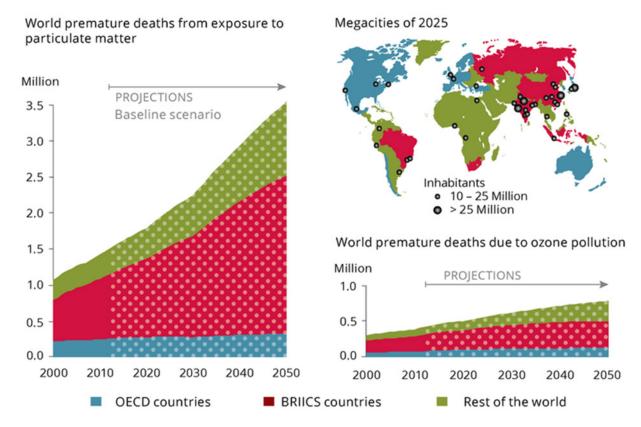


Figure 1. Projected premature deaths due to PM and ground-level ozone [7].

Exposure to high ambient PM concentrations does not only occur outside, in streets where the traffic intensity is high, but also in buildings due to ventilation. Since people roughly spend 85–90% of their time indoors [14], PM exposure is certainly a very serious problem, putting health at risk.

# Air purification in garages

Various solutions are available to limit ambient PM concentrations, such as porous media filters, cyclonic separators, wet scrubbers and electrostatic precipitators (ESP). ENS Technology [15], a manufacturer of ESPs based on a positive ionization technology, applied two of their ESP units in a parking garage in Cuijk (the Netherlands). The technology captures fine particles and ultrafine particles without the use of traditional filtering techniques. In the ESP, particles are bound together, forming coarse dust that cannot be inhaled (since it is not airborne). One of the advantages of this system is that most of the energy penalty, associated with the pressure drop across media filters (traditional technique), is eliminated. Furthermore, it is found that the applied systems do not add to the normal background ozone level, unlike several other ESP-based technologies. In Cuijk it is found that PM<sub>10</sub> concentrations are strongly reduced in the parking garage, but also in the shopping mall connected to the garage [16]. This leads to the idea of large-scale application of ESP units in parking garages, which would thus function as 'lungs of the city'.

### **CFD model Eindhoven city center**

A computational fluid dynamics (CFD) case study is conducted for the city center of Eindhoven (the Netherlands), including ESP units in 16 parking garages. Eindhoven is located in the south of the Netherlands and it is the fifth largest city of the Netherlands with 225,020 inhabitants [17]. The city center of Eindhoven is characterized by a mixture of low-rise and a few high-rise buildings, both commercial and residential, with open areas such as roads, parks and squares. The generated computational grid consisting out of 65.7 million cells, as presented in **Figure 2**, covers an area of about 5.1 km<sup>2</sup>.

To determine sources of  $PM_{10}$  in the model, traffic intensity, traffic emission data, parking garage use and background  $PM_{10}$  concentrations are considered. Terrain roughness is taken into account by assigning the correct aerodynamic roughness height  $(y_0)$ , equivalent sand-grain roughness height  $(k_S)$  and roughness constant  $(C_S)$  to the wall surfaces. Simulations are performed for a reference wind speed  $(U_{ref})$  of 1 m/s at 10 m height and a south-east wind direction. The southeast wind direction is chosen because ambient concentrations in Eindhoven are generally higher for this wind direction [18]. Thermal effects are not taken into account.

ESP units are implemented in the garages by means of a  $PM_{10}$  sink term. The units run at a volume flow rate of 9,000 m<sup>3</sup>/h and a  $PM_{10}$  removal efficiency of 70%

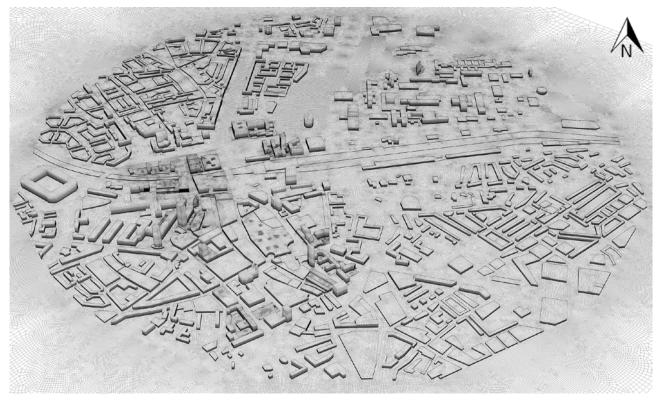
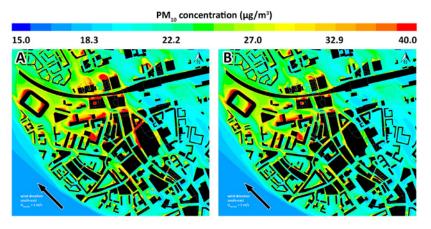
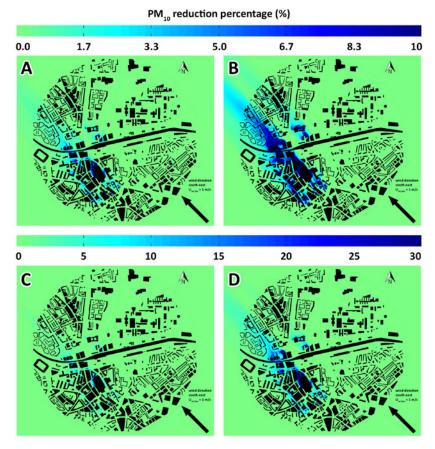


Figure 2. Overview of the computational grid of the city center of Eindhoven (65.7 million cells) [3].

[16]. Three different cases are modeled: a reference case without ESP units, a case including 99 units (one unit per 65 parking spots) and a case including 594 units (six units per 65 parking spots) spread over the 16 parking garages. Garage ventilation is imposed based on the Dutch building regulations (10.8 m³/m²h).



**Figure 3.** Contours of PM<sub>10</sub> concentrations ( $\mu$ g/m³) in a horizontal plane at 1.75 m height without units (a) and including 594 units (b) installed over 16 parking garages in the city center of Eindhoven. Wind direction South-East U<sub>ref,10m</sub> = 1 m/s.



**Figure 4.** Contours of PM<sub>10</sub> concentration reduction percentage at a height of 1.75 m for the case with 99 units (a, c) and 594 units (b, d) relative to the reference case without units. Wind direction South-East  $U_{ref,10m} = 1 \text{ m/s}$ . [3]

In this study, the steady Reynolds-averaged Navier-Stokes (RANS) approach is used in combination with the realizable k-ɛ turbulence model [19] for turbulence closure. A sub-configuration validation study is conducted using wind tunnel measurements of gas dispersion in regular arrays of rectangular building blocks by Garbero et al.

[20]. The reader is referred to Blocken et al. [3] for more information on this validation study.

#### Case study results

**Figure 3** presents the PM<sub>10</sub> concentration contours at a height of 1.75 m (pedestrian level) for the reference case without ESP units (**Figure 3a**) and for the case including 594 ESP units (**Figure 3b**). Results are presented for the region where most of the parking garages are located and where the urban density is relatively high. The concentrations exceed 60 μg/m<sup>3</sup> but in this case the maximum concentration shown is limited to 40 μg/m<sup>3</sup> for visualization purposes.

Figure 3 shows that relatively high PM<sub>10</sub> concentrations are found near the exhaust openings of the parking garages (> 40  $\mu$ g/m<sup>3</sup>). High PM<sub>10</sub> concentrations are found in the area with a high concentration of parking garages. Furthermore, in this area the urban density is relatively high compared to the rest of the city, resulting in relatively low wind velocities. **Figure 3** shows that PM<sub>10</sub> concentrations are significantly reduced near the garages and in the exhaust flow from the garage. This is also shown in **Figure 4** where the PM<sub>10</sub> concentration is compared with the reference case. Figure 4a, c presents the PM<sub>10</sub> reduction percentage for the case including 99 ESP units and Figure 4b, **d** presents the  $PM_{10}$  reduction for the case including 594 ESP units. For the case with 99 ESP units, reductions of up to 10% are found near the garage. However, further away from the garages the reductions are insignificant. For the case with 594 ESP units, reductions of up to 30% are found near the garages (locally even up to 50%). Further away from the garages reductions of up to 10% are found.

#### Discussion and conclusion

Although the study has some limitations, it can be concluded that local removal of PM in parking garages can be an effective strategy towards higher outdoor air quality. As previously mentioned, studies have shown that PM exposure is strongly related to human morbidity and mortality. For this reason the presented work could be very interesting for cities in which problems related to PM exposure are severe (e.g. where the urban density is high). Further research will focus on the limitations of the current study. First of all, the maximum PM<sub>10</sub> concentrations in the current simulations are rather limited (< 80 µg/m<sup>3</sup>). This is due to the fact that all the emissions are uniformly

distributed over the subdomains (instead of assigning them to the exact locations of streets) and the emissions are averaged over a period of 10 hours. It can be expected that during rush hours local concentrations are higher. For this reason unequal spreading, both in time and space, of traffic and thus of PM emissions will be taken into account. In addition, thermal effects can be included in future studies. In this way, the dispersion of emissions can be taken into account in a more accurate way and different meteorological situations can be analyzed. While the current research work and suggestions will certainly not solve the PM city problem completely, at least it can be a significant step in the right direction. ■

### Acknowledgements

Twan van Hooff is currently a postdoctoral fellow of the Research Foundation – Flanders (FWO) and is grateful for its financial support (project FWO 12R9715N). The authors also gratefully acknowledge the partnership with ANSYS CFD.

#### References

- [1] WHO 2014. Ambient (outdoor) air quality and health. Fact sheet N°313, updated March 2014. http://www.who.int/mediacentre/factsheets/fs313/en/
- [2] WHO 2013. Health effects of particulate matter policy implications for countries in eastern Europe, Caucasus and central Asia, ISBN 9789289000017.
- [3] Blocken B., Vervoort R., van Hooff T. (2016). Reduction of outdoor particulate matter concentrations by local removal in semi-enclosed parking garages: a preliminary case study for Eindhoven city center, J. Wind Eng. Ind. Aerodyn. 159, pp. 80-98.
- Chen H., Kwong J.C., Copes R., Tu K., Villeneuve P.J., van Donkelaar A., Hystad P., Martin R.V., Murray B.J., Jessiman B., Wilton A.S., Kopp A., Burnett R.T. (2017). Living near major roads and the incidence of dementia, Parkinson's disease, and multiple sclerosis: a population-based cohort study, Lancet 389, 718-726.
- [5] Rijksinstituut voor Volksgezondheid en Milieu (RIVM). Luchtkwaliteit.
- [6] OECD 2012. OECD Environmental Outlook to 2050: the Consequences of Inaction. 350 p. 9789264122246 (PDF).
- [7] EEA 2015. World premature deaths due to urban pollution from particulate matter and ground-level ozone. http://www.eea.europa.eu/ data-and-maps/figures/world-premature-deaths-due-to. European Environment Agency.
- WHO 2015. Economic costs of health impact of air pollution in Europe: clean air, health and wealth. http://www.euro.who.int/\_\_data/assets/pdf\_file/0004/276772/Economic-cost-health-impact-air-pollution-en.pdf?ua=1
- [9] WHO, 2005. Air quality guidelines, global update 2005. Particulate matter, ozone, nitrogen dioxide and sulfur dioxide. ISBN 92 890 2192 6.
- [10] WHO 2016. WHO Global Urban Ambient Air Pollution Database. http://www.who.int/phe/health\_topics/outdoorair/databases/cities/en/
- [11] Rogge WF, Hildemann LM, Mazurek MA, Cass GR, Simoneit BRT. 1993. Sources of fine organic aerosol. 3. Road dust, tire debris, and organometallic brake lining dust — roads as sources and sinks. Environm. Sci. & Techn. 27(9): 1892–1904.
- [12] Kim SR, Dominici F, Buckley TJ. 2007. Concentrations of vehicle-related air pollutants in an urban parking garage. Environ. Res. 105(3): 291–299.
- [13] Zhang GS, Li TT, Luo M, Liu JF, Liu ZR, Bai YH. 2008. Air pollution in the microenvironment of parked new cars. Build. Environ. 43(3): 315–319.
- [14] Valavanidis A, Fiotakis K, Vlacogianni T. 2008. Airborne particulate matter and human health: toxicological assessment and importance of the size and composition of particles for oxidative damage and carcinogenic mechanisms. J. Environ. Sci. Health Pt. C - Environ. Carcinog. Ecotoxicol. Rev. 26(4): 339-362.
- [15] ENS. 2008. Environmental Nano Solutions. www.enstechnology.nl/en.
- [16] ENS. 2016. Environmental Nano Solutions. Personal communication B. Blocken and ENS. 30 August 2016.
- [17] Eindhoven Municipality, 2015. http://www.eindhoven.nl/home.htm.
- [18] AiREAS. Meetdata Eindhoven, www.aireas.com.
- [19] Shih, T.H., Liou, W.W., Shabbir, A., Yang, Z., Zhu, J., 1995. A new k-e eddy viscosity model for high Reynolds number turbulent flows. Comput. Fluids 24, 227-238.
- [20] Garbero V, Salizzoni P, Soulhac L. 2010. Experimental study of pollutant dispersion within a network of streets. Bound.-Layer Meteor. 136: 45.